

93 DEC 13 AM 11:31

November 5, 1993

Ms. Susan Hugo  
Alameda County Department  
of Environmental Health  
80 Swan Way, Room 200  
Oakland, California 94621

Re: Greyhound Terminal  
Location 8934  
Oakland, California  
Preliminary Risk Evaluation

Dear Ms. Hugo:

On behalf of Greyhound Lines, Inc. (GLI), Engineering-Science, Inc. (ES) is pleased to provide the enclosed Preliminary Risk Evaluation report for the Greyhound Terminal in Oakland, California. This preliminary risk evaluation has been prepared in accordance with the discussions held in a meeting between GLI and the Alameda County Department of Environmental Health (ACDEH) on September 2, 1993. The risk assessment includes an evaluation of potential contaminant exposure pathways, existing contaminant levels and distribution, chemical characteristics, and site-specific factors such as soil permeability, and local land and water uses.

The results of this study indicate the lack of any significant health or environmental threat to current or future users of the site due to residual contaminants detected. Based on the results of this risk evaluation, alternative points of compliance, or site-specific cleanup criteria should be established. No soil remediation is recommended based on these results. Greyhound proposes to continue operation of the free product/groundwater treatment system on-site with monthly groundwater monitoring and system maintenance visits.

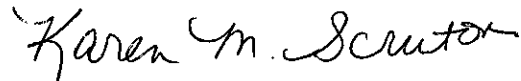
ENGINEERING-SCIENCE, INC.

Ms. Susan Hugo  
Alameda County Department of  
Environmental Health  
November 5, 1993  
Page 2

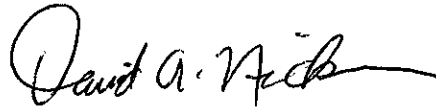
Greyhound looks forward to your favorable review of this report. In the interim, if you have any questions or require additional information, please contact us at (315) 451-9560.

Sincerely,

ENGINEERING-SCIENCE, INC.



Karen M. Scruton  
Project Toxicologist



David A. Nickerson  
Project Manager

KMS/DAN/lml

cc: T. Portele, GLI, Dallas, TX  
J.N. Baker, ES, Syracuse  
Richard Hiatt, RWQCB  
D. L. Chaffin, ES, Syracuse

**PRELIMINARY RISK EVALUATION**

**GREYHOUND TERMINAL**

**OAKLAND, CALIFORNIA**

Prepared for:

**GREYHOUND LINES, INC.  
DALLAS, TEXAS**

Prepared by:

**ENGINEERING-SCIENCE, INC  
290 ELWOOD DAVIS ROAD  
LIVERPOOL, NY 13088**

**OCTOBER, 1993**

## TABLE OF CONTENTS

EXECUTIVE SUMMARY .....	1
INTRODUCTION .....	2
SITE DESCRIPTION .....	2
Local Description and Surrounding Land Use and Climate .....	3
Local Geology .....	3
Hydrology.....	3
Local Surface Water and Groundwater Use.....	4
CONTAMINANTS OF CONCERN.....	4
Chemical and Physical Properties of the Chemicals of Concern.....	4
Contaminant Fate and Transport .....	5
Volatile Organic Compounds.....	5
Petroleum Hydrocarbons.....	6
ARAR IDENTIFICATION .....	6
Groundwater ARARs .....	7
Safe Drinking Water Act .....	7
Clean Water Act.....	7
Soil ARARs .....	8
PRELIMINARY HUMAN HEALTH EVALUATION .....	9
Exposure Assessment.....	9
Mechanisms of Migration .....	10
Pathways of Exposure.....	11
Toxicity Assessment .....	12
Health Criteria for Noncarcinogenic Effects.....	12
Health Criteria for Carcinogenic Effects .....	13
Toxicity Profiles For the Chemicals of Concern.....	15
Benzene .....	15
Ethylbenzene .....	16
Toluene.....	16
Xylenes.....	17
Petroleum Hydrocarbons.....	17
Risk Screening.....	18
Health and Environmental Criteria .....	18
Site Screening .....	18
Uncertainties in Risk Screening.....	18
PRELIMINARY ECOLOGICAL EVALUATION.....	19

CONCLUSIONS.....19  
REFERENCES.....20

## LIST OF TABLES

Table 1 Maximum Concentrations of Chemicals Analyzed

Table 2 Relevant Physical and Chemical Properties

Table 3 Applicable or Relevant and Appropriate Requirements -  
Groundwater

Table 4 Toxicity Values for the Evaluation of Potential  
Noncarcinogenic Effects

Table 5 Toxicity Values for the Assessment of Potential Carcinogenic  
Effects

Table 6 Comparison of Maximum Soil Contaminant Concentrations  
with Risk-Based PRGs

Table 7 Matrix of Potential Exposure Pathways

## LIST OF FIGURES

Figure 1 Location Map

Figure 2 Site Plan

**PRELIMINARY RISK EVALUATION  
GREYHOUND TERMINAL  
OAKLAND, CALIFORNIA**

**EXECUTIVE SUMMARY**

A preliminary risk evaluation was conducted for the Greyhound Terminal in Oakland, California. The purpose of the evaluation was to assess the potential risks to human health and the environment posed by contaminants detected at the facility and to determine the applicability of alternate site-specific cleanup levels. Site concentrations were compared to available or derived health and cleanup criteria.

For this preliminary risk assessment, the site was divided into two regions: the former Tank Pit area (source area) and the region surrounding the source area (perimeter). In the source area, concentrations of contaminants in groundwater exceeded criteria derived to protect both human health and the environment. A free-product/groundwater recovery system is currently in place in the source area. None of the chemicals detected in the groundwater in the perimeter were found to exceed the criteria used, indicating that the recovery system has been effective in preventing migration of contaminants from the source area.

Concentrations of BTEX in soils (benzene, toluene, ethylbenzene and xylenes) did not exceed calculated risk-based preliminary remediation goals (PRGs) in either the source area or the perimeter sample locations. Neither BTEX nor total petroleum hydrocarbons (TPH) were detected in the perimeter area, indicating that contamination has not migrated from the source area. Although TPH was detected in soils in the source area, risk-based PRGs could not be derived for these contaminants because USEPA-derived toxicity values are not available.

Currently, there are no individuals that may be exposed to either contaminated soil or groundwater, especially in the source area. The site is currently covered entirely by pavement and buildings, thereby eliminating contact with soils. The drinking water supply for the City of Oakland is currently municipally supplied with water originating in the Sierra Nevada Mountains, located 200 miles east of the site. Groundwater in the area is not considered a primary source of water supply because of the limited areal extent and thickness of the water-bearing unit. Consequently, it is highly unlikely that future users of the site will be exposed to groundwater. In the future, direct contact with soil is the only possible exposure pathway. Individuals may contact soil during future excavation and development of the site. As discussed above, however, neither BTEX in soil from the source area nor the perimeter were detected in concentrations above calculated risk-based cleanup levels. Given that the risks associated with TPH cannot be quantitated because USEPA-derived toxicity values are not available, it was concluded that a more detailed quantitative risk assessment is not appropriate.



## INTRODUCTION

The preliminary risk evaluation provides an assessment of the potential risks to human health and the environment posed by any contaminants detected during the site characterization and subsequent analyses. The data used in this evaluation were those collected for the 1989 Phase I Investigation performed by Brown and Caldwell (Brown and Caldwell, 1989), the 1992 Preliminary Site Investigation Report (ES, 1992) and the 1993 Supplemental Site Investigation Report (ES, 1993). The objectives of the preliminary risk evaluation are:

- To provide qualitative information on the potential risks to human and environmental receptors due to the release or threat of release of hazardous substances; and
- To help determine whether or not further action is required and the applicability of alternate, site-specific cleanup levels. Further action might include the collection of additional data and the preparation of a quantitative risk assessment.

The results of the site investigations were used to identify 1) the affected environmental media, 2) the contaminants of concern, 3) the chemical concentrations present, and 4) the potentially-exposed human or environmental receptors.

The initial steps of the preliminary evaluation were to:

- Summarize the available chemical sampling data;
- Review the factors that affect migration of contaminants, and identify and evaluate potential migration pathways;
- Evaluate the potential toxicities associated with exposure to the selected chemicals by human or environmental receptors; and
- Identify potential hazards to human or environmental receptors that may be affected by the migration of contaminants along identified pathways.

The site characterization is the initial investigation of a site and its purpose is to confirm or deny the presence of contamination. In keeping with the preliminary nature of the investigation, detailed calculations to quantify risk to human health and the environment from the site were not performed. Instead a qualitative approach was taken in which all potential receptors and exposure pathways were evaluated as to the possibility of pathway completion. Completed pathways were further evaluated using various cleanup criteria. These criteria may be existing standards, calculated health-based criteria, or existing cleanup levels. This evaluation was based on existing site information concerning migration pathways, the location and types of contaminants present, and the location of current and possible future receptors. Conservative assumptions were employed to ensure that potential exposure pathways (current and future) were not excluded from consideration.

## **SITE DESCRIPTION**

This section presents a description of the site, including surrounding land use, climate, geology, hydrology and surface and groundwater use.

### **Local Description and Surrounding Land Use and Climate**

The triangular-shaped Greyhound site is located at the corner of San Pablo Avenue and Castro Street near the central business district of Oakland, California (Section 26, Township 1S, Range 4W, United States Geological Survey, 1980)(Figure 1 and 2). The site, which encompasses an area of approximately 61,250 square feet, is entirely paved with asphalt and concrete. Land use in the immediate vicinity of the terminal is mixed (commercial/residential). The terminal is bordered on the west by Castro Street, on the east by San Pablo Avenue and Martin Luther King Way, on the south by 20th Street and on the north by Castro Street/San Pablo Avenue.

The Oakland area has a climate characterized by mild wet winters and warm dry summers. Precipitation is seasonal, falling mostly between November and March. Average annual precipitation recorded over a ninety year period (1885-1975) for the Oakland area is 20 inches (Alameda County Flood Control and Water Conservation District, 1988).

### **Local Geology**

The site is located in the San Francisco Bay Region of the Coast Ranges Geomorphic Province of California (Norris and Webb, 1990). The area, including the Greyhound facility, is underlain by a thick sequence of unconsolidated Pleistocene deposits that include the Merritt Sand and Older Alluvium. The Merritt Sand encountered directly below the site consists of loose, moderate-to well-sorted, fine- to medium-grained, clayey to silty sand and lenses of sandy clay and clay. It has a maximum thickness of 65 feet in the East Bay Plain Area. Underlying the Merritt Sand is 700 to 800 feet of Older Alluvium comprised of poorly consolidated to unconsolidated clay, silt, sand and gravel (Helley, Lajoie and Burke, 1972; Alameda County Flood Control and Water Conservation District, 1988).

Beneath the Greyhound facility, the subsurface materials encountered during previous investigations include sand, silt and clay. The predominant materials encountered during boring operations were silty, sandy clay with layers/lenses of fine-to medium-grained sand, silty sand and clayey sand interbedded within layers of clay and silty clay. Hydraulic conductivity values reported for silty, sandy clays range from  $10^{-9}$  to  $10^{-4}$  cm/sec (Freeze and Cherry, 1979; Domenico and Schwartz, 1990).

### **Hydrology**

The nearest surface water body to the Greyhound site is Lake Merritt (Figure 1), located approximately 1,700 feet east of the site. Lake Merritt is a brackish-water estuarine environment, connected to and influenced by the tidal fluctuations of San Francisco Bay. The Oakland Inner Harbor, the closest portion of the bay, is located approximately 2,700 feet south-southwest of the site.

The Greyhound facility lies within the Merritt Sand subarea of the East Bay Plain groundwater basin. Locally, groundwater is encountered under water-table conditions at a depth of 18 to 22 feet below ground surface (bgs). Regional groundwater flow is to the west-southwest (Alameda County Flood Control and Water Conservation District, 1988). At the Greyhound site, shallow groundwater flow is to the southeast.

#### **Local Surface Water and Groundwater Use**

The City of Oakland obtains its municipal and industrial water from the East Bay Municipal Utility District (EBMUD). EBMUD imports this water primarily from the surface waters of the Sierra Nevada Mountain Range, located approximately 200 miles east of the site.

Groundwater in the area is utilized for limited irrigation and industrial purposes. The area is not considered a primary source of water supply because of the limited areal extent and thickness of the water-bearing unit (Alameda County Flood Control and Water Conservation District, 1988).

Approximately 384 wells are located within Section 26, Township 1S, Range 4W (ACPWA, 1991). The vast majority (99%) of these wells are used to monitor or extract contaminated groundwater at commercial/industrial sites. One of the wells is used to supply water for irrigation. None of the wells located in Section 26 are used for the municipal water supply.

#### **CONTAMINANTS OF CONCERN**

Five parameters were assessed in groundwater and soil samples from the site: benzene, ethylbenzene, toluene, and xylenes (BTEX), and total petroleum hydrocarbons (TPH). During this and previous site investigations, TPH was further characterized as TPH from diesel (TPHD) or from gasoline (TPHG) (ES, 1992; ES, 1993).

For this preliminary risk assessment, the site was divided into two distinct portions: the source area (Tank Pit area) and the area surrounding the source area (perimeter). The source area is defined as those groundwater and soil samples taken from within and immediately around the former Tank Pit area. These include the following monitoring well locations: BC-1 through BC-3 from the 1989 investigation (Brown and Caldwell, 1989), and ES-1 through ES-5 from the 1992 investigation (ES, 1992). The area surrounding the source area is characterized as those groundwater and soil samples around the perimeter of the site, which were taken during the 1993 investigation (ES-6 through ES-11) (ES, 1993).

The contaminants detected in the groundwater and soil samples in both of these areas, and their maximum concentrations are listed in Table 1. The maximum concentrations for each chemical provide an indication of the magnitude of chemical contamination in the analyzed samples. The maximum values are used in subsequent sections to estimate the potential for adverse health affects. This approach will greatly overestimate any risks.

## **Chemical and Physical Properties of the Chemicals of Concern**

Physical and chemical properties of the chemicals of concern will affect fate and transport of those chemicals in the environment. Table 2 summarizes several important physical and chemical properties for the selected chemicals of concern.

The water solubility of a substance is a critical property affecting environmental fate. Highly-soluble chemicals can be rapidly leached from wastes and soils and are generally mobile in groundwater. Solubilities can range from less than 1 mg/L to totally miscible, with most common organic chemicals falling between 1 mg/L and 1,000,000 mg/L (Lyman et al., 1982). The water solubility of chemicals may become enhanced in the presence of organic solvents (e.g., toluene), which themselves are more soluble in water.

Volatilization of a compound will depend on its vapor pressure, water solubility, and air diffusion coefficient. Highly water-soluble compounds generally have lower volatilization rates from water unless they also have high vapor pressures. Vapor pressure, a relative measure of the volatility of chemicals in their pure state, ranges from roughly 0.001 to 760 millimeters of mercury (mm Hg) for liquids. The Henry's Law Constant, which combines vapor pressure with solubility, is more appropriate than vapor pressure alone for estimating releases from water to air for compounds having Henry's Law Constants. Compounds with Henry's Law Constants greater than  $10^{-3}$  atmospheres - cubic meter per mole (atm-m<sup>3</sup>/mole) may readily volatilize from water if not bound to organic matter; those with values ranging from  $10^{-3}$  to  $10^{-5}$  are associated with moderate volatilization, while compounds with values less than  $10^{-5}$  will only volatilize from water to a limited extent (Lyman et al., 1982).

The organic carbon partition coefficient ( $K_{oc}$ ) reflects the propensity of a compound to sorb to organic matter found in soil. The normal range of  $K_{oc}$  values is 1 to  $10^7$  milliliters per gram (mL/g), with higher values indicating greater sorption potential. Chemicals which have a strong tendency to sorb to organic matter (i.e., chemicals with high  $K_{oc}$  values) will move more slowly in the environment than chemicals with low  $K_{oc}$  values.

### **Contaminant Fate and Transport**

The previous subsection provided a description of the important physical and chemical properties of the contaminants detected at the site. This subsection describes how these properties affect the persistence and transport of these compounds in the environment.

#### **Volatile Organic Compounds**

Benzene, ethylbenzene, toluene, and xylenes are all volatile organic compounds. These compounds have high Henry's Law Constants, moderate to high solubilities and low  $K_{oc}$  values. This indicates that these compounds can be expected to be mobile in the environment.

All four of the volatile compounds have been detected in groundwater on site, and all, except benzene, have been detected in soil samples. The properties that enhance the mobilities of these compounds also make them more available for

degradation. Because of their high vapor pressures these compounds would be expected to volatilize from uncovered surface soils. However, volatilization of these compounds from site subsurface soils would be inhibited by the overlying pavement and building foundations.

### **Petroleum Hydrocarbons**

During the 1992 investigation (ES, 1992), an analysis of TPH as diesel was performed and, during the 1993 investigation (ES, 1993), TPH analyses for both diesel (TPHD) and gasoline (TPHG) were completed based on a requirement of the Alameda County Department of Environmental Health (ACDEH).

TPHD and TPHG concentrations were measured in accordance with DHS/LUFT Method (EPA Method 3510/8015 for TPHD and EPA Method 8015M for TPHG).

The fate of petroleum hydrocarbons in soils is affected primarily by their distribution, volatility, and leaching potential. Low molecular weight aromatic hydrocarbons, such as BTEX, partially evaporate. The remaining hydrocarbons will migrate to different depths in the soil column and possibly to groundwater.

The aliphatic organics which represent the residual compounds have negligible water solubilities, low vapor pressures and high adsorption coefficients. The proportion of petroleum hydrocarbons that will adsorb to soil particles rather than continue migration depends on the type of soil, the particular petroleum product involved, the volume of the release, and the amount of rainfall. In general, leaching to groundwater is favored by high rainfall and permeable soils. Leaching potential also increases for petroleum compounds with high solubility and low adsorption coefficients.

Most compounds measured as petroleum hydrocarbons are relatively persistent in the environment. Biodegradation is the main elimination mechanism, but natural rates are somewhat slow, especially for cyclic or aromatic hydrocarbons. Complete natural biodegradation of petroleum hydrocarbons may require many years (API, 1986).

### **ARAR IDENTIFICATION**

In evaluating the degree of contamination at a site, consideration must be given to applicable or relevant and appropriate requirements (ARARs) of Federal and State environmental laws. Applicable requirements are those cleanup standards, standards of control, and other substantive environmental protection requirements, criteria, or limitations that specifically address a hazardous substance, contaminant, remedial action, or other circumstance at a site. Relevant and appropriate requirements are those standards, standards of control, and other substantive environmental protection requirements, criteria, or limitations that, while not "applicable", address problems or situations sufficiently similar to those encountered at the site that their use is well suited to the particular site.

Many local implementing agencies, such as ACDEH, use the LUFT Manual (1989) as guidance in the assessment and cleanup of UFST sites. Three site

investigation categories are recognized by the LUFT manual: Category 1, no evidence of soil contamination; Category 2, known soil contamination; and Category 3, known or suspected groundwater contamination. Analytical results for soil and groundwater samples collected during the 1989 investigation (Brown and Caldwell Consulting Engineers, 1989), the preliminary site investigation (ES, 1992) and the supplemental site investigation (ES, 1993) indicate that the site falls under Category 3 (known or suspected groundwater contamination).

### **Groundwater ARARs**

Potential ARARs for water quality at the site include the Safe Drinking Water Act, and the Clean Water Act. Standards and criteria promulgated under these programs are provided in Table 3 for potential contaminants of concern at the site. California Water Quality Standards for the compounds of concern are also provided on Table 3.

#### **Safe Drinking Water Act**

The Safe Drinking Water Act (SDWA) mandates USEPA to establish regulations to protect human health from contaminants in drinking water. USEPA has promulgated drinking water standards which generally apply to community water systems. Primary drinking water standards include maximum contaminant levels (MCLs) and maximum contaminant level goals (MCLGs). MCLs are set at levels that are protective of human health, while taking into account available treatment technologies and the costs to large public water systems. MCLGs are strictly health-based and do not take cost or feasibility into account. Secondary drinking water regulations consist primarily of secondary maximum contaminant levels (SMCLs) for specific contaminants or water characteristics that may affect the aesthetic qualities of drinking water (i.e., color, odor, and taste).

USEPA MCLs and MCLGs for contaminants of concern at the site are provided in Table 3 (USEPA, 1993a). California MCLs are also identified (USEPA, 1992). The levels indicated are potential ARARs. Maximum site concentrations for the source area and the perimeter area are also presented in Table 3.

In the source area, the USEPA and/or California MCLs were exceeded for benzene, ethylbenzene, toluene and xylenes. MCLs are not available for TPHD (maximum concentration of 950 mg/L) or TPHG (maximum concentration of 1.5 mg/L). Given that it is highly unlikely that the groundwater at the site would ever be used as a source of drinking water, BTEX and TPHD/TPHG in the groundwater should not pose a hazard to either current or future receptors.

In the perimeter area, none of the USEPA and/or California MCLs were exceeded for benzene, ethylbenzene, toluene or xylenes. Neither TPHD nor TPHG were detected in any of the perimeter wells.

#### **Clean Water Act**

The Clean Water Act (CWA) requires the establishment of guidelines and standards to control the direct or indirect discharge of pollutants to waters of the

United States. The standards required by the CWA include water quality criteria for specific pollutants. USEPA has developed two kinds of water quality criteria: one for the protection of human health and another for the protection of aquatic life. These criteria are non-enforceable guidelines used by the states to set water quality standards for surface water. These non-enforceable standards are potential ARARs when the state has not promulgated water quality standards for the specific pollutants and water bodies of concern.

Table 3 provides the water quality criteria for the potential contaminants of concern at the site (USEPA, 1986a). Although these values are not applicable, they may be relevant and appropriate because groundwater at the site may eventually discharge to surface water (presumably Lake Merritt). Given that the surface water in the vicinity is brackish, water quality criteria for marine aquatic life are reported since these values are more conservative than those for freshwater aquatic life. Data, however, are insufficient to derive criteria for the marine life, therefore, the values presented in the table represent the Lowest-Observed-Effect-Level (LOEL). Given that the contaminants are expected to be diluted and degraded prior to reaching surface water in the area, it is highly unlikely that any of the site chemicals would ever reach any surface water body at concentrations above detection limits.

In the source area, acute and/or chronic ambient water quality criteria for marine aquatic life were exceeded for benzene and ethylbenzene. Criteria were not exceeded for toluene. Criteria were not available for xylenes and TPHs. It should be reiterated, however, that the values presented as criteria for these chemicals are LOEL values since data were insufficient for the development of criteria.

In the perimeter area, water quality criteria were not exceeded for benzene, ethylbenzene or toluene. Criteria are not available for xylenes and TPHs.

#### **Soil ARARs**

Exposure via soil is a possible pathway of concern for future site users if excavation is undertaken at the site. In California, soil cleanup levels at sites with underground storage tanks are evaluated on a case-by-case basis using guidance contained in the LUFT guidance (1989) and the Water Quality Control Plan of the San Francisco Bay Region (1992). The LUFT guidance focuses on the cleanup of soils to levels that will not adversely affect groundwater. The LUFT guidance evaluates three types of site-specific data in their general risk appraisal: precipitation, depth to groundwater and extent of soil contamination.

Given the conditions at the Greyhound site, the LUFT guidance indicates that soils should be cleaned up to non-detectable levels for the protection of groundwater. This is not appropriate for the Greyhound site, however, since the groundwater is currently being treated with a groundwater recovery system, which has been shown to be effective in preventing the migration of contamination from the source area. At this site, it is more appropriate to implement the guidance stated in the Water Quality Control Plan of the San Francisco Bay Region (1992). This guidance states that, if it is unreasonable to cleanup soils to background concentration levels, soil cleanup levels may be derived under the following conditions:

- allow residual pollutants to remain in soil at concentrations such that:
  - (a) any leachate generated would not cause ground water to exceed applicable ground water quality objectives, and
  - (b) health risks from surface and subsurface exposure are within acceptable guidelines.
- require follow-up groundwater monitoring to verify that groundwater is not polluted by chemicals remaining in the soil. Follow-up groundwater monitoring may not be required where residual soil pollutants are not expected to impact groundwater.
- require measures to ensure that soils with residual pollutants are covered and managed to minimize pollution of surface waters and/or exposure to the public.
- where significant amount of wastes remain onsite, implement institutional controls to the extent applicable. This may include, but is not limited to, subsurface barriers, pollutant immobilization, toxicity reduction, financial assurances.

Given that a groundwater recovery system is currently in place at the site, it is reasonable to derive soil cleanup levels that provide for the protection of human health following direct soil contact. USEPA (1991) provides guidance for the derivation of Preliminary Remediation Goals (PRGs) which represent initial cleanup goals. The risk-based PRGs were derived using a USEPA-defined conservative exposure scenario for soil ingestion and appropriate USEPA-derived toxicity values (discussed below under Toxicity Assessment; see Table 4 and Table 5). This scenario is based on a 30-year, time-weighted exposure of an individual ingesting soil. The exposure was time-weighted to account for differences in body weight and ingestion rate between children and adults (child: 6 year exposure of a 15 kg child to 200 mg soil/day; adult: 24 year exposure of a 70 kg adult to 100 mg soil/day). The exposure was assumed to occur 350 days/year and was averaged over 30 years for noncarcinogenic effects and 70 years for carcinogenic effects.

Maximum detected site concentrations of contaminants and the appropriate risk-based PRGs are presented in Table 6. Risk-based PRGs were not exceeded for BTEX in either the source area or in the perimeter area. Risk-based PRGs could not be calculated for TPHD, TPHG or TFH because USEPA has not derived toxicity values for these contaminants.

## **PRELIMINARY HUMAN HEALTH EVALUATION**

The previous subsections have described the physical and chemical properties of contaminants found at the site and the effect these properties will have on environmental fate and transport mechanisms as well as identifying ARARs that can be used to evaluate the significance of the sampling results. This subsection provides the preliminary human health evaluation which includes an assessment of exposure routes, a description of contaminant toxicities and presents the initial human health risk screening.



## Exposure Assessment

The presence of a contaminant in a particular environmental medium does not necessarily indicate that human exposure will occur. In order for human exposure to occur, a complete exposure pathway must exist. A complete exposure pathway consists of the following:

- A contaminant source and mechanism for release;
- An environmental transport medium;
- An exposure point; and
- A human receptor and a feasible route of exposure at the exposure point.

If any of the items listed above are missing then an exposure pathway is incomplete. The following paragraphs describe the transport mechanism and exposure pathways for the site.

### Mechanisms of Migration

The media into which a contaminant migrates affects the types of human and environmental exposures which may occur. The previous subsections have described the physical and chemical properties of concern. This subsection discusses the mechanisms of contaminant migration and potential exposure routes for the site. Contaminants have been detected in groundwater and soil samples at the site. Several mechanisms exist through which contaminants may migrate.

**Migration into Air.** Contaminants may migrate into the air through three primary mechanisms: volatilization, soil gas migration, and suspension of soil particles (wind erosion or mechanical disturbances).

Volatilization is the mass transfer of a compound from a specific medium (such as soil) to the air. Environmental factors that affect volatilization include temperature, soil porosity, soil water content, soil organic carbon content, and depth of contamination (Jury et al., 1983). Volatilization may be an important migration pathway for contaminants having high vapor pressures (greater than 100 mm Hg) or high Henry's Law Constants (greater than  $10^{-3}$  atm-m<sup>3</sup>/mole). The contaminants detected at the site which are in this category are benzene, ethylbenzene, toluene and xylenes. Volatilization from soils, however, is unlikely to occur given that the site is completely covered by pavement and building foundations.

Fugitive dust emissions from wind or vehicle disturbances at the site are unlikely to occur now or in the future. The site is completely covered by pavement and building foundations.

**Leaching into Groundwater.** The percolation of rainwater through the soils at the site is unlikely, since the site is paved. Contamination of groundwater at the site by BTEX, TPHD and TPHG has been detected in the source area.

Once contaminants reach the groundwater, the individual chemicals will have different migration rates. Volatile organic compounds were detected in both the soil and groundwater in the source area. Given the nature of the soil at the site, (silty, sandy clay), soil permeability is expected to be low to moderate. The

presence of pavement and the building foundations can be expected to slow contaminant migration off site. The fact that significant contamination was not found in the perimeter area indicates that contaminants are not migrating from the source area.

**Migration in Surface Water.** Contaminant migration into surface water may occur through groundwater discharge downgradient of the site. The regional flow pattern is southeast, towards Lake Merritt. Significant groundwater contamination has not been detected in the area immediately outside of the source area, indicating that contaminant migration is impeded by the low permeability soils and that the groundwater recovery system in place at the site is effective in containing the contaminants. Consequently, the amount of site contaminants reaching surface water should be below detection limits.

### **Pathways of Exposure**

Potential pathways of exposure to chemicals of concern at the site have been summarized in Table 7. Demographic and land use information were used in developing exposure pathways. The USEPA requires that hypothetical future use of a site be considered as well as current use. Therefore, this table presents current-use and future-use pathways.

Current-use pathways include exposure to air, groundwater, surface water, and soils by local residents and on-site workers. Because the site is entirely covered by pavement and buildings, direct contact with soil and the generation of airborne volatiles from soil will be negligible. This indicates that the potential for completion of the soil and air pathways is very low.

The primary source of drinking water for the City of Oakland is surface waters of the Sierra Nevada Mountain Range, located approximately 200 miles east of the site. Groundwater in the area is utilized for limited irrigation and industrial purposes. The area is not considered a primary source of water supply because of the limited areal extent and thickness of the water-bearing unit (Alameda County Flood Control and Water Conservation District, 1988). Consequently, the groundwater pathway is incomplete for current nearby residents.

Exposure to contaminated surface water resulting from migration of chemicals off-site is highly unlikely since the amounts of contamination that would reach this water would be negligible. Groundwater in the perimeter of the source area did not contain significant contamination, indicating that the groundwater recovery system currently in place at the site is effective in containing contamination from the source area. The surface water pathway, therefore, is incomplete.

Future-use pathways include exposure to contaminated air, groundwater, surface water, and soils by hypothetical future residents and construction workers. The probability of future-use pathway completion is higher than that for current use, based primarily on the remote possibility of future development of the site for residential or commercial use. By the time development might occur on the site, contaminant concentrations in groundwater will have diminished significantly from current levels as a result of the installed groundwater recovery system, and

concentrations in soil and water will have also diminished due to natural processes involving degradation and/or dispersion.

Similar to current-use pathways, exposure to contaminants through volatilization and fugitive dust generation is expected to be minimal. Exposure to contaminated groundwater through its development as a potable drinking water source is unlikely because future residents, if any, would be supplied with water by the City of Oakland. Exposure to contaminated surface water in the future is highly unlikely, as the amounts of contamination that would reach this water would be negligible. Exposure of hypothetical future residents and construction workers to contamination in soils is unlikely, but will be further assessed.

### **Toxicity Assessment**

The objective of the toxicity assessment is to weigh available evidence regarding the potential for particular contaminants to cause adverse effects in exposed individuals and to provide, where possible, an estimate of the relationship between the extent of exposure to a contaminant and the increased likelihood and/or severity of adverse effects. The types of toxicity information applied in this assessment include the reference dose (RfD) used to evaluate noncarcinogenic effects and the slope factor used to evaluate carcinogenic potential.

#### **Health Criteria for Noncarcinogenic Effects**

For chemicals that exhibit noncarcinogenic effects, many authorities consider organisms to have repair and detoxication capabilities that must be exceeded by some critical concentration (threshold) before the health effect is manifested. For example, an organ can have a large number of cells performing the same or similar functions that must be significantly depleted before the effect on the organ is seen. This threshold view holds that a range of exposures from just above zero to some finite value can be tolerated by the organism without an appreciable risk of adverse effects.

Health Criteria for chemicals exhibiting noncarcinogenic effects for use in risk assessment are generally developed using USEPA RfDs (USEPA, 1993b,c). In general, the RfD is an estimate of route-specific average daily intake (dose) for individuals (including sensitive individuals) below which there will not be an appreciable risk of adverse health effects. The RfD is derived using conservative safety factors (e.g., to adjust from animals to humans and to protect sensitive subpopulations) to ensure that it is unlikely to underestimate the potential for adverse noncarcinogenic effects to occur. The purpose of the RfD is to provide a benchmark against which the sum of other doses (e.g., those projected from human exposure to various environmental conditions) might be compared. Doses that are significantly higher than the RfD may indicate an inadequate margin of safety could exist for exposure to that substance and that an adverse health effect could occur.

USEPA has developed oral and inhalation RfDs for most of the chemicals of concern selected for the site. In addition, the chemicals of concern may affect different target organs in the body. Some of the chemicals of concern that may have

noncarcinogenic effects following long-term exposure, and the target organs that are most sensitive to these chemicals, are as follows:

- Ethylbenzene and xylenes may adversely affect the liver.
- Ethylbenzene may adversely affect the kidney.
- Toluene and xylenes may adversely affect the nervous system.
- Total xylenes may cause reproductive effects.

Potential noncarcinogenic effects with RfDs of chemicals of concern identified at the site are shown on Table 4.

No RfDs or slope factors are available for the dermal route of exposure. In some cases, however, noncarcinogenic or carcinogenic risks associated with dermal exposure can be evaluated using an oral RfD or an oral slope factor. Exposures via the dermal route generally are calculated and expressed as absorbed doses. These absorbed doses are compared to an oral toxicity value that is also expressed as an absorbed dose. This requires quantitative estimates of both oral and dermal exposure and is only appropriate for chemicals causing systemic toxicity. However, since the purpose of this evaluation is to provide only a gross estimate of site hazards, exposures via the dermal route will not be further considered.

#### **Health Criteria for Carcinogenic Effects**

For chemicals that exhibit carcinogenic effects, most authorities recognize that one or more molecular events can evoke changes in a single cell or a small number of cells that can lead to tumor formation. This is the nonthreshold theory of carcinogenesis which purports that any level of exposure to a carcinogen can result in some finite possibility of generating the disease. Generally, regulatory agencies assume the non-threshold hypothesis for carcinogens in the absence of information concerning the mechanisms of action for the chemical.

USEPA's Carcinogen Assessment Group (CAG) has developed slope factors (i.e., dose-response values) for estimating excess lifetime cancer risks associated with various levels of lifetime exposure to potential human carcinogens. The carcinogenic slope factors can be used to estimate the lifetime excess cancer risk associated with exposure to a potential carcinogen (USEPA, 1993b,c). Risks estimated using slope factors are considered unlikely to underestimate actual risks, but they may overestimate actual risks. Excess lifetime cancer risks are generally expressed in scientific notation and are probabilities. An excess lifetime cancer risk of  $1 \times 10^{-6}$  represents the probability that one individual out of one million will develop cancer as a result of exposure to a carcinogenic chemical over a 70-year lifetime under specified exposure conditions. USEPA has suggested developing remedial alternatives for cleanup of Superfund sites using total excess lifetime cancer risks ranging from  $10^{-4}$  to  $10^{-6}$ . Potential carcinogenic targets with associated slope factors for the chemicals of concern identified at the site are shown on Table 5.

In practice, slope factors are derived from the results of human epidemiology studies or chronic animal bioassays. The data from animal studies are generally

fitted to the linearized multistage model and the dose-response curve is obtained. The 95th percentile upper confidence limit slope of the dose-response curve is subjected to various adjustments and an interspecies scaling factor is applied to conservatively derive the slope factor for humans. Thus, the actual risks associated with exposure to a potential carcinogen quantitatively evaluated based on animal data are not likely to exceed the risks estimated using these slope factors, but they may be much lower. Dose-response data derived from human epidemiological studies are fitted to dose-time-response curves on an ad hoc basis. These models provide rough but plausible estimates of the upper limits on lifetime risk. Slope factors based on human epidemiological data are also derived using very conservative assumptions and, as such, they too are considered unlikely to underestimate risks. In summary, while the actual risks associated with exposures to potential carcinogens are unlikely to be higher than the risks calculated using a slope factor, they could be considerably lower.

In addition, there are varying degrees of confidence in the weight of evidence for carcinogenicity of a given chemical. USEPA has proposed a system for characterizing the overall weight of evidence for a chemical's carcinogenicity based on the availability of animal, human, and other supportive data (USEPA, 1986b). The weight-of-evidence classification is an attempt to determine the likelihood that an agent is a human carcinogen and thus qualitatively affects the estimation of potential health risks. Three major factors are considered in characterizing the overall weight of evidence for carcinogenicity: (1) the quality of evidence from human studies, (2) the quality of evidence from animal studies, which are combined into a characterization of the overall weight of evidence for human carcinogenicity; and (3) other supportive information which is assessed to determine whether the overall weight of evidence should be modified. USEPA's final classification of the overall weight of evidence includes the following five categories:

- **Group A - - Human Carcinogen**

This category indicates that there is sufficient evidence from epidemiological studies to support a causal association between an agent and cancer.

- **Group B - - Probable Human Carcinogen**

This category generally indicates that there is at least limited evidence from epidemiological studies of carcinogenicity to humans (Group B1) or that, in the absence of adequate data on humans, there is sufficient evidence of carcinogenicity in animals (Group B2).

- **Group C - - Possible Human Carcinogen**

This category indicates that there is limited evidence of carcinogenicity in animals in the absence of data on humans.

- **Group D - - Not Classified**

This category indicates that the evidence for carcinogenicity in animals is inadequate.

- **Group E - - No Evidence of Carcinogenicity to Humans**

This category indicates that there is no evidence for carcinogenicity in at least two adequate animal tests in different species, or in both epidemiological and animal studies.

Slope factors are developed based on epidemiological or animal bioassay data for a specific route of exposure, either oral or inhalation. For some chemicals, sufficient data are available to develop route-specific slope factors for inhalation and ingestion. For chemicals with only one route-specific slope factor but for which carcinogenic effects may also occur via another route, USEPA has used the available value to evaluate risks associated with both potential routes of exposure. Slope factors, classifications, and potential carcinogenic effects for the chemicals of concern identified at the site are shown in Table 5.

### **Toxicity Profiles For the Chemicals of Concern**

Environmental Fate and Toxicity information for both human and environmental receptors is included in the following paragraphs.

#### **Benzene**

Benzene is a colorless aromatic hydrocarbon with a characteristic odor. Benzene was widely used in the past as a solvent and as an octane-raising additive in gasoline. Presently, benzene is used primarily in the chemical industry where it is used as a starting or intermediate material for the synthesis of many other organic compounds (ATSDR, 1989a).

Benzene has been shown to be mobile in the soil/groundwater system. It is relatively soluble in groundwater and may be transported through sandy soils and soils of low organic content. The amount of benzene sorbed to the soil increases with increasing organic content. Benzene is highly volatile, and volatilization in surficial soils is probably an important transport mechanism, however, sorption of benzene vapors onto soil particles may slow the vapor-phase transport. Hydrolysis is not expected to be an important process for benzene degradation. In soils, benzene is biodegraded both aerobically and anaerobically by microorganisms. The specific organisms which biodegrade benzene, however, are found in small numbers in soil (ATSDR, 1989a).

Ambient water quality criteria for benzene were developed by USEPA for acute exposure of freshwater organisms and for both acute and chronic exposure of marine organisms (Table 3; USEPA, 1986a). No information is available on the toxicity of benzene to terrestrial wildlife, domestic animals, birds, or plants. Toxic effects in laboratory animals include central nervous system effects, hematological effects, and immune system depression (ATSDR, 1989a).

Benzene is readily absorbed following oral and inhalation exposure (ATSDR, 1989a). The toxic effects of benzene in humans following exposure by inhalation is the same as that for laboratory animals and includes central nervous system effects, hematological effects, and immune system depression. In humans, acute exposure to high concentrations of benzene vapors has been associated with dizziness, nausea, vomiting, headache, drowsiness, narcosis, coma, and death (Sittig, 1985). Chronic

exposure to benzene vapors can produce reduced leukocyte, platelet, and red blood cell levels (ATSDR, 1989a).

Benzene is a human carcinogen (USEPA Group A). Chronic exposure to benzene is known to cause leukemia and bone marrow damage. In addition, the compound is a central nervous system depressant at high concentrations, and may cause acute narcotic reactions (Sittig, 1985).

#### **Ethylbenzene**

Ethylbenzene is a colorless aromatic liquid. It is used in industry as a resin solvent and in the conversion to styrene monomer (ATSDR, 1990a). No empirical studies on the bioaccumulation of ethylbenzene were found. No information on the toxicity of ethylbenzene to terrestrial wildlife or birds was available. The USEPA has developed ambient water quality criteria for ethylbenzene for acute exposure of freshwater and marine organisms (Table 3; USEPA, 1986a).

Ethylbenzene is moderately adsorbed to soils and will leach to groundwater, particularly in soils containing low levels of organic matter. Ethylbenzene will volatilize from surface soils and will be biodegraded by microbial populations (ATSDR, 1990a).

Ethylbenzene has been classified as a Group D carcinogen, indicating that there is no evidence that ethylbenzene causes cancer in humans or animals. In humans, short-term inhalation exposure to ethylbenzene can result in sleepiness, fatigue, headache, and mild eye and respiratory irritation. The liver, kidneys and developing fetus appear to be the primary targets following chronic oral exposure (ATSDR, 1990a).

#### **Toluene**

Toluene is a monocyclic, aromatic, colorless liquid used as a component of gasoline. It is used in manufacturing benzene, urethane foams and other organic compounds. Toluene functions as a solvent in products such as cleaning agents (ASTDR, 1989b).

From the available data, it appears that volatilization is the major route of removal from surface water and soils. Toluene is rapidly degraded in air, soil and water. Toluene will be adsorbed by sediments and suspended solids, especially those rich in organic matter (ATSDR, 1989b).

The USEPA has developed ambient water quality criteria for acute exposure of freshwater organisms and for both acute and chronic exposure of marine organisms (Table 3; USEPA, 1986a). Toluene is a volatile compound that has been found to readily move from water to the atmosphere. Toluene has a moderate tendency to bioaccumulate in the fatty tissues of aquatic species (ATSDR, 1989b).

Little information is available on the toxicity of toluene to terrestrial species. Information on the toxicity of toluene to plants or avian species is not available.

Toluene is absorbed in humans following all routes of exposure (ATSDR, 1989b). Toluene has been classified as a Group D carcinogen, indicating that there is no evidence that toluene causes cancer in humans or animals. In humans, the

primary acute effects of toluene vapors are central nervous system depression and narcosis. Also seen at low levels of exposure are irritation of the skin and eyes and impairment of coordination and reaction time when inhaled. In humans, chronic exposure to toluene vapors has been associated with central and peripheral nervous system effects, hepatomegaly, and hepatic and renal function changes. Effects on the liver and central nervous system have also been observed in animals following oral exposure (ATSDR, 1989b).

### **Xylenes**

Xylenes are mixtures of the ortho, meta, and para isomers, with the meta form usually the principal component. Xylenes may also contain impurities such as benzene, trimethylbenzene, toluene, phenol, thiophene and pyridene. The xylenes are widely used as fuel components and as solvents for paints and coatings. Xylenes are commonly used in the chemical industry as intermediates. Specifically, ortho-xylene is used in the manufacture of phthalic anhydride, which is a basic building block for plasticizers. Meta-xylene is an intermediate in the preparation of isophthalic acid, which is the base of unsaturated polyester resins. Commercially, para-xylene is the most important isomer, most of which converted to terephthalic acid or dimethylterephthalate and used to make ppolyester resins (ATSDR, 1990b).

Xylenes are relatively mobile in soil/water systems, especially in aqueous phase. Volatilization is the primary transport mechanism for xylenes. Xylenes will slowly biodegrade in subsurface soils (ATSDR, 1990b).

The xylenes have been classified as Group D carcinogens, indicating that there is no evidence that the xylenes cause cancer in humans or animals. The three xylene isomers have similar toxicological properties and are discussed together. Approximately 60 percent of an inhaled dose is absorbed, and absorption of orally-administered xylenes is nearly complete. Dermal absorption is reported to be minor following exposure to xylene vapor, but may be significant following contact with the liquid (ATSDR, 1990b)). Liquid xylenes and high vapor concentrations are irritating to the eyes and skin and may result in severe respiratory and central nervous system effects. Symptoms include dizziness, drowsiness, nausea, vomiting, abdominal pain, loss of appetite, pulmonary edema, and unconsciousness, as well as reversible effects on the liver and kidneys. The effects of chronic exposure resemble those from acute exposure, but are more severe (ATSDR, 1990b).

### **Petroleum Hydrocarbons**

Petroleum hydrocarbons are a group of compounds that are thick, dark yellow to brown, or green-black liquids which consist of a mixture of hydrocarbons from  $C_2H_2$  and up. They are used as a source of gasoline, petro ether, petrolatum, fuel and lubricating oils, butane, and isopropyl alcohol. The petroleum hydrocarbons found in diesel fuel are the ones of primary concern in this risk evaluation.

Hydrocarbon-containing petroleum residues are decomposed in soil systems. Hydrocarbons degrade to carbon dioxide and water via several intermediates (organic acids, ketones, aldehydes, alcohols, and other hydrocarbon derivatives). Nonvolatile components of oils tend to stay tightly bound in soil, while volatile



fractions may escape into the atmosphere. No significant loss or movement of oil through surface runoff from rainfall or downward leaching occurs.

Fuel oil is a combustible liquid and a skin irritant. Breathing oil mists may irritate the nose and throat. Chronic exposure to oil mists may lead to the development of lipoid pneumonia. Similarly refined and processed petroleum residual materials have been shown to cause skin cancer and liver damage in laboratory animals through prolonged skin contact. There is no direct evidence that fuel oil causes skin cancer or liver damage in humans.

## **Risk Screening**

### **Health and Environmental Criteria**

In this section concentrations of chemicals of concern are compared with appropriate criteria to provide a rough estimation as to whether the contaminants pose a risk. The method is intended as a preliminary screening tool rather than a detailed evaluation of risks posed by contaminants at the site. Where ARARs are not developed, other information may be needed to determine what is protective of human health and the environment. Other criteria to be used for comparison purposes include health-based levels derived from toxicity data, and published target cleanup levels.

Current information on the health and environmental effects of various toxicants, including slope factors and RfDs were obtained from the Integrated Risk Information System (IRIS; USEPA, 1993b) and the Health Effects Assessment Summary Tables (HEAST; USEPA, 1993c). IRIS is a computerized library of current information that is updated on a continuous basis. It contains health risk assessment information on chemicals that have undergone a detailed review of toxicity data by work groups composed of USEPA scientists from several agency program offices, and represents an USEPA consensus. Information includes RfDs and slope factors for systemic toxicants. These values are used to calculate human health-based PRGs according to USEPA guidelines.

### **Site Screening**

Exposure via soil is a possible pathway of concern for future site users if excavation is undertaken. A comparison of concentrations of chemicals detected in soil at the site with calculated PRGs is presented in Table 6. The contaminants in site subsurface soils did not exceed any of the calculated PRGs.

Exposure via inhalation of volatile organic contaminants or suspended particulates contaminated with volatile organic contaminants is unlikely due to low concentrations of these compounds in surface soils and the presence of overlying pavement and building foundations.

### **Uncertainties in Risk Screening**

In quantifying risks for a given receptor who is exposed to multiple chemicals by a number of different pathways, USEPA generally assumes that the total risk incurred by the receptor is essentially a sum of the individual risks incurred by each chemical and pathway of exposure. This is reflected in the USEPA methodology

used to quantify both noncarcinogenic and carcinogenic risks. Thus, the potential for adverse effects in a given receptor will increase with the number of chemicals detected at the site and the number of pathways by which the receptor could be exposed.

The preliminary risk assessment methodology employed here uses Health Criteria which were developed as though each chemical in question were present at the site by itself and as though exposure would occur through only one pathway (oral exposure to the medium of concern). Thus the risk screening process employed in this assessment considers only oral exposure pathways and does not apportion risks among the total number of chemicals detected at the site. Although in most cases, oral exposure will account for most contaminants, dermal exposure may also be of concern. If risks were apportioned among the total number of chemicals present and all possible pathways of exposure, Health Criteria values would be lower.

Negative findings in the screening process do not guarantee that there are no potential risks. If a baseline risk assessment were conducted in accordance with USEPA guidance documents (USEPA, 1989, 1990, 1991), the total risks for each receptor would be summed across all chemicals and pathways of exposure. Thus it is possible that unacceptable risks could be calculated in a baseline assessment even though the preliminary risk screening indicated that potential adverse health effects were not likely to occur. A negative risk screen thus should not be used as the sole basis for eliminating a site from further investigation or concern.

#### **PRELIMINARY ECOLOGICAL EVALUATION**

The area surrounding the site is typical of an urban area which has been significantly impacted by many anthropogenic sources over many years. There is little or no habitat of any significance within several miles of the site. The surrounding area consists of lands occupied by streets, parking lots, buildings, and other structures. The only wildlife expected to be in the immediate area includes that normally found in urban environments such as rodents (rats and mice) and birds (pigeons, starlings, and sparrows), and any wildlife present would not be exposed to site contaminants. Therefore, risks to nearby flora and fauna will not be further evaluated.

#### **CONCLUSIONS**

This preliminary risk evaluation does not indicate the presence of any significant health or environmental threat to current or future users of the site due to the contaminants detected at the Greyhound Oakland Terminal. The only potential exposure pathway identified was ingestion of soils by hypothetical future residents and workers. PRGs were derived for all chemicals of concern, except for TPHs because USEPA-derived toxicity values are not available for this group of compounds. Consequently, the potential toxicity of TPHs to future receptors could not be screened. Maximum site concentrations of BTEX were below the calculated PRG levels.

A groundwater recovery system and monitoring program are already in place. The recovery system appears to be effective in that contamination at the site is restricted to the source area: no significant contamination was reported in the perimeter area of the site. It is recommended that the recovery system and monitoring be continued in the near future and in the long-term, be replaced by groundwater monitoring. Additionally, no soil corrective action is recommended at this time. Site-specific cleanup levels which would allow existing residual soil contamination to be left in place are justified by the results of this risk assessment.

Given that toxicity values are not available for TPH, and that all other chemicals of concern were below calculated PRGs, a more detailed quantitative risk assessment is not appropriate.

## REFERENCES

- ACPWA, 1991. Alameda County Public Works Agency. Bay Plain Groundwater Study. Well Inventory Report. October 7, 1991.
- Alameda County Flood Control and Water Conservation District. 1988. Geohydrology and Groundwater Quality Overview of the East Bay Plain Area. Alameda County, California. 205(j) Report. June, 1988.
- API, 1986. The Migration of Petroleum Products in Soil and Groundwater: Principles and Countermeasures. American Petroleum Institute. Publication No. 4149.
- ATSDR, 1989a. Toxicological Profile for Benzene. Draft. Agency for Toxic substances and Disease Registry. May, 1989.
- ATSDR, 1989b. Toxicological Profile for Toluene. Draft. Agency for Toxic substances and Disease Registry. December, 1989.
- ATSDR, 1990a. Toxicological Profile for Ethylbenzene. Draft. Agency for Toxic substances and Disease Registry. December, 1990.
- ATSDR, 1990b. Toxicological Profile for Xylenes. Draft. Agency for Toxic substances and Disease Registry. December, 1990.
- Brown and Caldwell Consulting Engineers, 1989. Phase I Investigation. Oakland California Terminal. Consultant's report prepared for Greyhound Lines, Inc. Dallas, Texas. June, 1989.
- Domenico, P.A. and F.W. Schwartz, 1990. Physical and Chemical Hydrogeology. New York, NY. John Wiley and Sons, Inc. 824 p.
- ES, 1992. Preliminary Site Investigation Report, Greyhound Terminal, Oakland, California. Prepared for Greyhound Lines, Inc. Dallas, Texas. January 1992.
- ES, 1993. Supplemental Site Investigation Report, Greyhound Terminal, Oakland, California. Prepared for Greyhound Lines, Inc. Dallas, Texas. October, 1993.
- Freeze, R.A. and J.A. Cherry, 1979. Groundwater. Engelwood Cliffs, New Jersey. Prentice Hall, Inc. 604 p.
- Helly, E., Lajoie, K.R. and D.B. Burke, 1972. Geologic Map of Late Cenozoic Deposits. Alameda County, California. United States Geological Survey. MF 429.
- Howard, P.H., 1989. Handbook of Environmental Fate and Exposure Data For Organic Chemicals, Vol. I: Large Production and Priority Pollutants. Lewis Publishers, Inc. Chelsea, Michigan. 574 pp.
- Jury, W.A., Spencer, W.F. and W.J. Farmer, 1983. Behavior Assessment Model for Trace Organics in Soil, I. Model Description. J. Environ. Qual. Vol. 12, No. 4.

- LUFT Manual, 1989. State of California Leaking Underground fuel Tank Task Force. Leaking Underground Fuel Tank Manual: Guidelines for Site Assessment. Cleanup and Underground Storage Tank Closure. October, 1989.
- Lyman, W.J., Rechl, W.G. and D.H. Rosenblatt, 1982. Handbook of Chemical Property Estimation Methods: Environmental Behavior of Organic Compounds. McGraw Hill Book Company. New York.
- Norris, R. and R. Webb, 1990. Geology of California. 2nd Edition. John Wiley and Sons, Inc. 541 p.
- Sax, N.I. and R.J. Lewis, 1989. Dangerous Properties of Industrial Materials. 7th ed., Vol. I. Van Nostrand Reinhold. New York.
- Sittig, M., 1985. Handbook of Toxic and Hazardous Chemicals and Carcinogens. 2nd Edition. Noyes Publications. Park Ridge, New Jersey. 950 pp.
- USEPA, 1986a. Quality Criteria for Water 1986. Office of Water Regulations and Standards. Washington, DC EPA 440/5-86-001.
- USEPA, 1986b. Guidelines for Carcinogenic Risk Assessment. U.S. Environmental Protection Agency. Federal Register: Vol 51, No 185. September 24, 1986.
- USEPA, 1991. Risk Assessment Guidance for Superfund: Volume I - Human Health Evaluation Manual Part B. Development of Risk-based Preliminary Remediation Goals. Interim. Office of Emergency and Remedial Response. Publication 9285.7-01B. December 1991.
- USEPA, 1992. Environmental Reporter. Bureau of National Affairs, Inc. Update October 23, 1992.
- USEPA, 1993a. Drinking Water Regulations and Health Advisories. U.S. Environmental Protection Agency. Office of Water. May, 1993.
- USEPA, 1993b. Integrated Risk Information System (IRIS). Office of Research and Development. Online database. September, 1993.
- USEPA, 1993c. Health Effects Assessment Summary Tables (HEAST). Office of Emergency and Remedial Response. OHEA ECAO-CIN-909. March, 1993.
- United States Geological Survey, 1980. Topographic Map - Oakland West, 7.5 Minute Quadrangle.
- Water Quality Control Plan of the San Francisco Bay Region Update. 1992. Ground Water Basin Plan Amendments. Final: Adopted October 21, 1992.

**TABLE 1**  
**MAXIMUM CONTAMINANT CONCENTRATIONS**  
**GREYHOUND TERMINAL**  
**OAKLAND, CALIFORNIA**

<u>Contaminant</u>	<u>Maximum Concentration</u> (ppm)	
	<u>Source</u>	<u>Perimeter</u>
<u>GROUNDWATER</u>		
Benzene	2.1	ND
Ethylbenzene	0.84	ND
Toluene	3.9	0.0007
Xylenes	6.0	0.0012
TPHD	950	ND
TPHG	1.5	ND
<u>SOILS</u>		
Benzene	ND	ND
Ethylbenzene	27	ND
Toluene	49.5	ND
Xylenes	150	ND
TPHD	160	ND
TPHG	NA	ND
TFH	4260	NA

ND = Parameter analyzed but not detected above the method detection limit.

NA = Parameter not analyzed.

TABLE 2  
RELEVANT PHYSICAL AND CHEMICAL PROPERTIES (1)

CONTAMINANT	CAS NUMBER	WATER SOLUBILITY (mg/l) at 25°C	VAPOR PRESSURE (mm Hg) at 25°C	HENRY'S LAW CONSTANT (atm-m <sup>3</sup> /mole)	K <sub>oc</sub> (ml/g)
BENZENE	71-43-2	1,787.0	9.52E+01	5.50E-03	.3-100
ETHYLBENZENE	100-41-4	160.0	9.53E+00	7.90E-03	165
TOLUENE	108-88-3	515 (20°C)	22 (20°C)	6.60E-03	295
XYLENES (MIXED)	1330-20-7	130	6.72 (21°C)	5.2E-3 - 7.6E-3	47.7-260
TPH	-	-	-	-	-

(1) Source: ATSDR 1989a,b; 1990a,b

TABLE 3  
GROUNDWATER ARARs  
GREYHOUND TERMINAL  
OAKLAND, CALIFORNIA

CONTAMINANT	MAXIMUM GROUNDWATER CONCENTRATIONS (mg/L)		CALIFORNIA MCLs (1) (mg/L)	SAFE DRINKING WATER ACT (mg/L) (2)		FEDERAL AMBIENT WATER QUALITY CRITERIA (mg/L) (3)			
	SOURCE	PERIMETER		MCL	MCLG	Human Health Consumption		Marine Aquatic Life (a)	
						Water and Fish Ingestion	Fish Ingestion Only	Acute Criteria	Chronic Criteria
Benzene	2.1	ND	0.001	0.005	0	0.00066 (b)	0.040 (b)	5.1 (c)	0.7 (c)
Ethylbenzene	0.84	ND	0.68	0.7	0.7	1.4	3.28	.43 (c)	—
Toluene	3.9	0.0007	—	1.0	1.0	14.3	424.0	6.3 (c)	5.0 (c)
Xylenes	6.0	0.0012	1.75	10.0	10.0	—	—	—	—
TPHD	950	ND	—	—	—	—	—	—	—
TPHG	1.5	ND	—	—	—	—	—	—	—

Footnotes:

- (a) Given that most of the surface water in the vicinity of the site is brackish, values for marine life were used since they are more conservative than those for fresh water.
- (b) Criteria based on carcinogenicity (10E-6 risk).
- (c) Insufficient data were available to develop criteria, the values presented are LOELs (Lowest-Observed-Effect-Levels).

References:

- (1) Environmental Reported. Bureau of National Affairs, Inc. Update 10/23/92 (USEPA, 1992).
- (2) Drinking Water Regulations and Health Advisories. USEPA: Office of Water. May, 1993 (USEPA, 1993a).
- (3) Quality Criteria for Water. USEPA: Office of Water Regulations and Standards. May, 1986 (USEPA, 1986a).



TABLE 4  
TOXICITY VALUES FOR THE EVALUATION OF  
POTENTIAL NONCARCINOGENIC EFFECTS (1)

CONTAMINANT	CHRONIC INHALATION RfC (mg/cu. m)	CHRONIC ORAL RfD (mg/kg/day)	CRITICAL EFFECT	
			INHALATION	ORAL
BENZENE	ND	ND	ND	ND
ETHYLBENZENE	1.0E+00	1.0E-01	Developmental toxicity	Liver, kidney toxicity
TOLUENE	4.0E-01	2.0E-01	CNS effects	Liver, kidney weights
XYLENES(TOTAL)	ND	2.0E+00	ND	Body Weight, Mortality, Hyperactivity
TPH	ND	ND	ND	ND

(1) Source: USEPA Integrated Risk Information System. Online. September 1993.  
CNS = Central Nervous System  
ND = No Data

**TABLE 5**  
**TOXICITY VALUES FOR THE ASSESSMENT OF**  
**POTENTIAL CARCINOGENIC EFFECTS (1)**

CONTAMINANT	SLOPE FACTOR 1/(mg/kg/day)		WEIGHT-OF-EVIDENCE CLASSIFICATION (2)		TUMOR SITE	
	INHALATION	ORAL	INHALATION	ORAL	INHALATION	ORAL
BENZENE	2.9E-02	2.9E-02	A	A	Blood Cells	Blood Cells
ETHYLBENZENE	NA	NA	D	D	NA	NA
TOLUENE	NA	NA	D	D	NA	NA
XYLENES (TOTAL)	NA	NA	D	D	NA	NA
TPH	NA	NA	NA	NA	NA	NA

(1) Source: USEPA Integrated Risk Information System (IRIS). Online. September, 1993.

(2) The Weight-of-Evidence Classification is defined in the text.

NA = Not Applicable.

TABLE 6  
 COMPARISON OF MAXIMUM SOIL CONTAMINANT  
 CONCENTRATIONS WITH RISK-BASED CLEANUP GOALS  
 GREYHOUND TERMINAL  
 OAKLAND, CALIFORNIA

CONTAMINANT	MAXIMUM CONCENTRATION (mg/kg)		RISK-BASED PRELIMINARY REMEDIATION GOALS (mg/kg) (1)	CONCENTRATION EXCEEDS CRITERION ?
	<u>Source</u>	<u>Perimeter</u>		
BENZENE	ND	ND	22	NO
ETHYLBENZENE	27	ND	27,000	NO
TOLUENE	49.5	ND	54,000	NO
XYLENE (TOTAL)	150	ND	540,000	NO
TPHD	160	ND	-	-
TPHG	NA	ND	-	-
TFH	4260	ND	-	-

(1) Preliminary Remediation Goals (PRGs) derived using USEPA methodology (USEPA, 1991).  
 ND = Not Detected; NA = Not Analyzed; (-) = Not Applicable

**TABLE 7**  
**MATRIX OF POTENTIAL EXPOSURE PATHWAYS**  
**GREYHOUND TERMINAL**  
**OAKLAND, CALIFORNIA**

Potentially Exposed Population	Exposure Route, Medium and Exposure Point	Pathway Selected for Evaluation ?	Reason for Selection or Exclusion
<u>Current Land Use</u>			
Local Residents	Ingestion of, dermal contact with, or inhalation of volatiles from down-gradient groundwater	No	There are no known private or industrial water wells that are used for drinking water. The City of Oakland is served by municipal supplies originating 200 miles east of the site.
Local Residents and Workers	Inhalation of volatiles at the site	No	Contaminated soils are at depth and are covered by pavement and the building foundations.
Local Residents	Ingestion of or dermal contact with affected soils at the site	No	Contaminated soils are at depth and are covered by pavement and the building foundations.

**TABLE 7 (continued)**  
**MATRIX OF POTENTIAL EXPOSURE PATHWAYS**  
**GREYHOUND TERMINAL**  
**OAKLAND, CALIFORNIA**

Potentially Exposed Population	Exposure Route, Medium and Exposure Point	Pathway Selected for Evaluation ?	Reason for Selection or Exclusion
<u>Future Land Use</u>			
Hypothetical Future Residents	Ingestion of, dermal contact with, or inhalation of volatiles from shallow downgradient groundwater	No	It is highly unlikely that shallow groundwater from the vicinity of the site would ever be used as a drinking water source. Future drinking water needs would be met via the municipal supply.
Hypothetical Future Residents and Construction Workers	Ingestion of or dermal contact with affected soils at the site	Yes	If excavation is undertaken on site, it is possible that workers could be exposed. The area is zoned commercial/residential, so it is possible that future residents could live on-site.

FIGURE 1

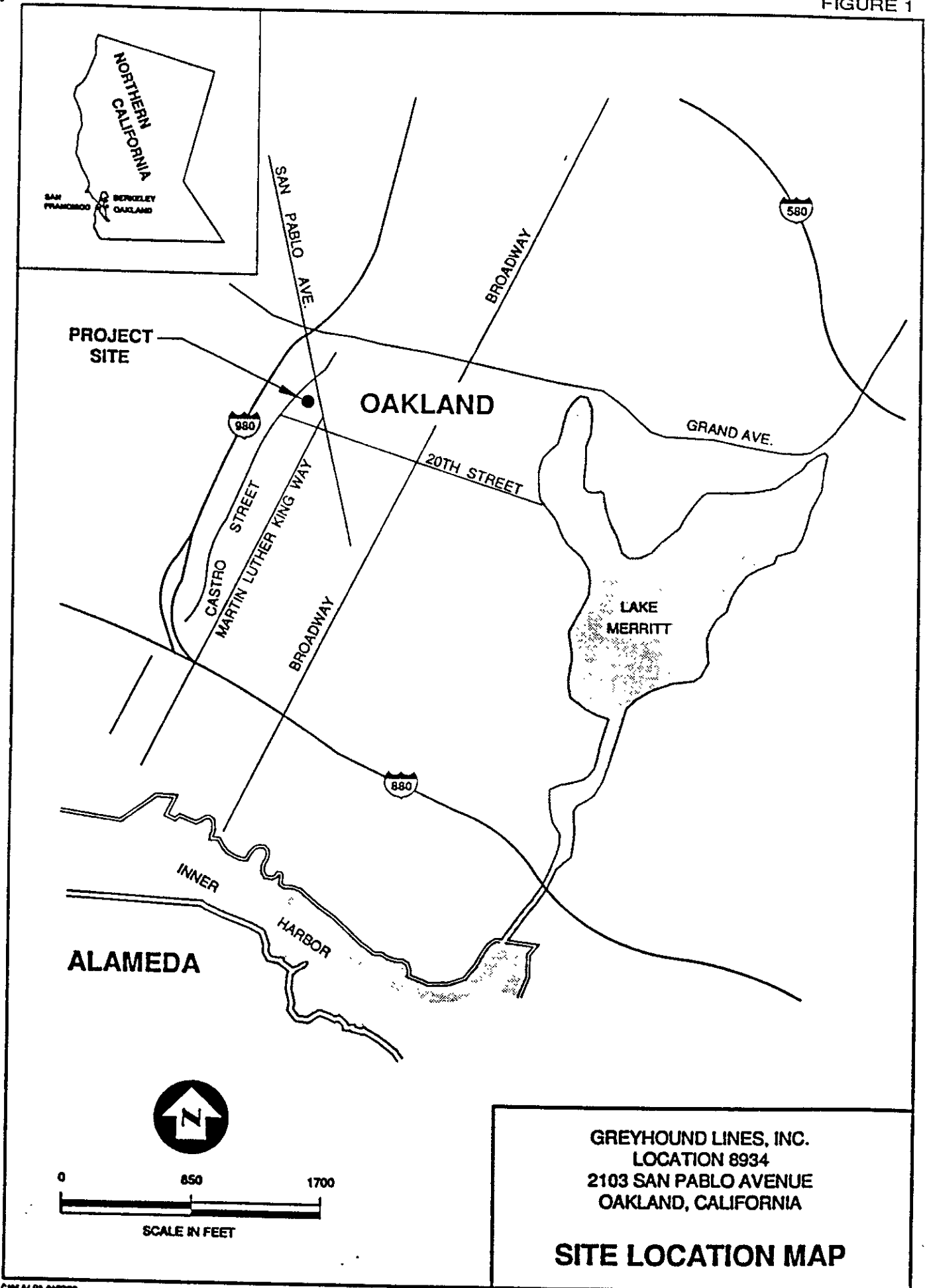


FIGURE 2

